Supplement of Web Ecol., 19, 15–26, 2019 https://doi.org/10.5194/we-19-15-2019-supplement © Author(s) 2019. This work is distributed under the Creative Commons Attribution 4.0 License.





Supplement of

Keep your enemies closer: enhancing biological control through individual movement rules to retain natural enemies inside the field

Thomas Delattre et al.

Correspondence to: Thomas Delattre (thomas.delattre@inra.fr)

The copyright of individual parts of the supplement might differ from the CC BY 4.0 License.

Additional information

S1 Presentation of the models following the ODD protocol

S1.1 Foraging model

S1.1.1 Model description

The description of our individual-based Foraging model follows the ODD protocol (Grimm et al., 2006, 2010). The model was implemented in NetLogo (Tisue and Wilensky, 2004).

S1.1.2 Purpose

The purpose of this model is to understand how the qualities of three categories of habitats in agricultural landscapes affect the residence time of natural enemies in the agricultural plot, and thereby affect the potential biological control. The Foraging model also serves as a "null hypothesis" by providing a most simplistic approach to movement rules that can be compared to more complex models.

S1.1.3 Entities, state variables, and scales

We are not representing pests in the model but only a generic natural enemy species.

The Foraging Model has only one entity, namely individual natural enemies. They are described by a set of simple state variables characterising the location of the natural enemy and its movement ability.

- Localisation (x, y)
- Habitat sensitivity (%)
- Movement ability (energy e)

Time steps (ts) are abstract, as well as space units (pixels). Space is described in two dimensions. The typical simulated plot is a 500 pixels wide square, but its size can be varied by the experimenter.

The three habitat types are the agricultural crops, the grassy field margins (GFMs) and the hedgerows. The quality of each habitat type can be varied so that each can be considered hostile, favourable or of intermediate quality, from the point of view of the natural enemy.

S1.1.4 Process overview and scheduling

The processes of the simulation model are described in the flowchart (Fig. S1).

At each time step, the submodel *Forage* is executed for all natural enemies in a random order, and defines their next location. The residence time calculations are then executed for each pixel, and summarised at the end of the simulation.

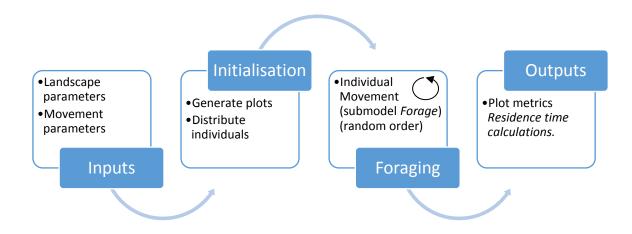


Figure S1.1. Flowchart of the Foraging model.

S1.1.5 Design Concepts

Basic principles: The movements of natural enemies are mainly foraging movements based on "movement ecology" and "habitat selection" literature and based on a simple non-specific behavioural assumption: movement is a biased random walk affected by local habitat quality (Bartumeus et al., 2005; Bell, 1991). As a result, movements in the model are in part imposed by the random walk, but some adaptation has been taken into account.

Emergence: The dynamic of natural enemies' movements and the resulting residence time in the habitat categories emerge from the foraging behaviour of the individuals. The interplay between movement, habitat qualities and their spatial organisation is not straightforward.

Adaptation: Natural enemies adapt their movement to the habitat cell they are to move on: better cells have a higher probability to be chosen for the next movement. A habitat sensitivity parameter is provided in the inputs, that increase the probability that the best cell is ignored, and a random cell is chosen instead. Due to these parameters, individual optimise the time spent foraging in favourable habitats, and minimise the time spent in unfavourable habitats.

Objectives: Not relevant. Learning: Not relevant. Interaction: Not relevant. Prediction: Not relevant.

Sensing: Natural enemies perceive the habitat quality of the cell they are on at the beginning of the time step, and that of the eight neighbouring cells.

Stochasticity: In the model, the construction of plots (Initialisation section f.) and the individual movements are stochastic. Movements are classically modelled by random processes (Codling et al., 2008) because unpredictability of food distribution for a predator implies stochasticity in the search.

Observation: At each time step, each pixel occupation status is stored, and its residence time is incremented if at least one individual is located on it. When the simulation is over (when all individuals have depleted their energy pool), the mean residence time and its variance, and the proportion of unvisited pixels are calculated for each habitat type, and stored for statistical calculations in R (R Core Team, 2011). Mean field residence time and its variance are calculated over the residence times of all pixels belonging to a given habitat type, summed over all simulation time steps.

S1.1.6 Forage Submodel

At each time step, individuals compare the habitat quality of eight neighbouring cells, and identify the best one. This core process may be affected, according to the *habitat sensitivity* parameter that has been introduced to compare different species responses to sets of habitat qualities. This parameter illustrates interspecific variability in sensitivity to habitat quality (*i.e.* generalist *vs* specialist species). The probability that a random cell is chosen instead of the better one is inversely proportional to the habitat sensitivity of the species (*i.e.* a species with a low habitat sensitivity would have a higher probability to ignore the better cells and engage in a pure random walk). At each time step, the remaining energy pool was decremented by 1-q/100, where q represents habitat quality of the current cell. This mechanism allows us to mimic the direct and indirect costs of movement (Bonte et al., 2012) that are high in hostile habitats and low in favourable habitats. The habitat sensitivity parameter (Table 1) is used to alter the effect of habitat quality on movement cost, as a proxy of interspecific differences in habitat sensitivity. A random value [-1 <RV> 1] is added to the pixel cost with a probability equal to the habitat sensitivity of the species/100 (i.e. adding noise around the cost value). The costs of diagonal and orthogonal moves are identical.

S1.1.7 Initialisation

The model is initialised by assigning habitat types to cells (either "hedgerow", "grassy field margin" or "agricultural plot"). A habitat quality parameter is then attributed to each pixel according to its habitat type, and the quality that has been attributed to it in the inputs. 2000 individuals are then distributed on random hedgerow cells, with a random initial orientation, and an energy pool of 500e representing their intrinsic initial movement ability.

S1.1.7.1 Plot Generation

Fields shapes and patterns are obtained using a method similar to a T-tessellation (Papaïx et al., 2014) that consists of seeding the landscape with a defined number of randomly distributed seeds, each of which is a departure point for three edges that eventually form a rectangle (Figure 1). This method allowed probabilistic control on the number of polygons, their size and shape, while exploring a diversity of spatial distributions of field shapes and sizes (Figure 1). In order to focus on habitat quality, the patch density is kept constant to maintain a stable landscape structure throughout the simulations (see Supplement S2 for the effect of patch density) and we alter only habitat quality for each landscape element (between extreme values 1 and 99, respectively hostile and favourable, other values ranging from 5 to 95 with a 5 interval). The landscape is a 500 pixels wide square treated as a torus, and is composed of 10 to 12 fields surrounded by 4 pixels-wide hedgerows and 5 pixels-wide GFMs (similar to a typical bocage landscape, Burel et al., 1998; Thenail and Baudry, 2004). Although the field-GFM-hedgerow trio is used as an example for clarity, the structure could apply to fields surrounded by other types of borders.

S1.1.7.2 Foraging Parameters

The values used in our case study for foraging parameters are provided in Table 1. They are designed to represent two hypothetical species, to illustrate the sensitivity of the model to differences in habitat sensitivity. An "insensitive species" with a movement behaviour that allows individuals to free themselves from local habitat conditions to reach more easily another region of the landscape: in the Foraging model that species is characterised by a lower value of the sensitivity to habitat quality parameter (Table1). On the contrary, the "sensitive species" is characterised by a movement behaviour that depended more strongly on local conditions (Table 1) with a higher sensitivity to habitat and lower directional persistence.

S1.2 The Routine & Direct Movements model

S1.2.1 Model description

The description of our individual-based Routine & Direct Movements (RDM) model follows the ODD protocol (Grimm et al., 2006, 2010). The model was implemented in NetLogo (Tisue and Wilensky, 2004).

S1.2.2 Purpose

The purpose of this model is to understand how the qualities of three categories of habitats in agricultural landscapes affect the residence time of natural enemies in the agricultural plot, and thereby affect the potential biological control. The RDM model is designed to illustrate a different approach to movement, compared to the Foraging model and the SMS. In the RDM model, individuals react to changes in habitat quality by changing the shape of their path and the probabilities to pass habitat boundaries (instead of choosing a destination cell at each step).

S1.2.3 Entities, state variables, and scales

We are not representing pests in the model but only a generic natural enemy species.

The RDM model has only one entity, namely individual natural enemies. They are described by a set of simple state variables characterising the location of the natural enemy and its movement ability.

- Localisation (x, y)
- Habitat sensitivity (%)
- Movement ability (energy e)

Time steps (ts) are abstract, as well as space units (pixels). Space is described in two dimensions. The typical simulated plot is a 500 pixels wide square, but its size can be varied by the experimenter.

The three habitat types are the agricultural crops, the grassy field margins (GFMs) and the hedgerows. The quality of each habitat type can be varied so that each can be considered hostile, favourable or of intermediate quality, from the point of view of the natural enemy.

S1.2.4 Process overview and scheduling

The processes of the simulation model are described in the flowchart (Fig. S1).

At each time step, the submodel *RDM* is executed for all natural enemies in a random order, and defines their next location. The residence time calculations are then executed for each pixel, and summarised at the end of the simulation.

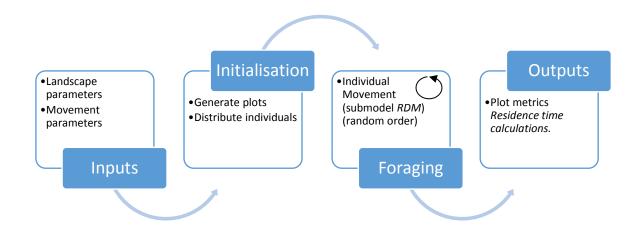


Figure S1.2. Flowchart of the RDM model.

S1.2.5 Design Concepts

Basic principles: The movements of natural enemies are mainly foraging movements based on "movement ecology" and "habitat selection" literature and based on simple non-specific behavioural assumptions: movement is a correlated random walk whose shape is affected by local habitat quality and contrast at habitat boundaries (Bartumeus et al., 2005; Bell, 1991; Van Dyck and Baguette, 2005). As a result, movements in the model are in part imposed by the random walk, but some adaptation has been taken into account.

Emergence: The dynamic of natural enemies' movement and the resulting residence time in the habitat categories emerge from the movement behaviour of individuals. The interplay between movement, habitat qualities and their spatial organisation is not straightforward.

Adaptation: Natural enemies adapt their movement to the habitat cell they are located on. On favourable habitat, they move slowly and sinuously, and tend to avoid crossing towards unfavourable habitats. On the contrary, on unfavourable habitats, they move fast and almost straight, and direct to each favourable habitat encountered. Due to these changes, they optimise the time they spend foraging in favourable habitats and minimise the time they spend in unfavourable habitats. The habitat sensitivity parameter is added to compare different scenarios with different species response to landscape. The effect of habitat quality on the sinuosity of the path and on the probability to cross a boundary are proportional to the habitat sensitivity of the species: an insensitive species will be more likely to ignore the current habitat quality when defining its path sinuosity, and to ignore the contrast of a boundary when deciding if it is to cross it.

Objectives: Not relevant.

Learning: Not relevant.

Interaction: Not relevant.

Prediction: Not relevant.

Sensing: Natural enemies perceive the habitat quality of the cell they are on at the beginning of the time step, and that of the eight neighbouring cells.

Stochasticity: In the model, the construction of plots (Initialisation section f.) and the individual movements are stochastic. Movements are classically modelled by random processes (Codling et al., 2008) because unpredictability of food distribution for a predator implies stochasticity in the search.

Observation: At each time step, each pixel occupation status is stored, and its residence time is incremented if at least one individual is located on it. When the simulation is over (when all individuals have depleted their energy pool), the mean field residence time and its variance, and the proportion of unvisited pixels are calculated for each habitat type, and stored for statistical calculations in R (R Core Team, 2011). Mean field residence time and its variance are calculated over the residence times of all pixels belonging to a given habitat type, summed over all simulation time steps.

S1.2.6 RDM Submodel

At each time step, individuals read the habitat quality of their current cell. According to their quality, they define the sinuosity of their path (a higher quality habitat causes higher sinuosity). The sinuosity of the path is then used to select stochastically a tentative cell for the next movement among the eight neighbour cells. If that tentative cell has a different habitat quality than the cell of origin, a boundary-crossing routine is executed. The individual chooses stochastically whether to cross that boundary, with a probability that is proportional to the contrast between both origin and destination pixel. The habitat sensitivity parameter was added in order to compare different scenarios with different species response to landscape. The effect of habitat quality on the sinuosity of the path and on the probability to cross a boundary are proportional to the habitat sensitivity of the species: an insensitive species will be more likely to ignore the current habitat quality when defining its path sinuosity, and to ignore the contrast of a boundary when deciding if it is to cross it.

At each time step, the remaining energy pool was decremented by 1-q/100, where q represents habitat quality of the current cell. This mechanism allows us to mimic the direct and indirect costs of movement (Bonte et al., 2012) that are high in hostile habitats and low in favourable habitats. The habitat sensitivity parameter (Table 1) is used to alter the effect of habitat quality on movement cost, as a proxy of interspecific differences in habitat sensitivity. A random value [-1 <RV> 1] is added to the pixel cost with a probability equal to the habitat sensitivity of the species/100 (i.e. adding noise around the cost value). The costs of diagonal and orthogonal moves are identical.

S1.2.7 Initialisation

The model is initialised by assigning habitat types to cells (either "hedgerow", "grassy field margin" or "agricultural plot"). A habitat quality parameter is then attributed to each pixel according to its habitat type, and the quality that has been attributed to it in the inputs. 2000 individuals are then distributed on a random hedgerow cell, with a random initial orientation, and an energy pool of 500e representing their intrinsic initial movement ability.

S1.2.7.1 Plot Generation

Fields shapes and patterns are obtained using a method similar to a T-tessellation (Papaïx et al., 2014) that consists of seeding the landscape with a defined number of randomly distributed seeds, each of which is a departure point for three edges that eventually form a rectangle (Figure 1). This method allowed probabilistic control on the number of polygons, their size and shape, while exploring a diversity of spatial distributions of field shapes and sizes (Figure 1). In order to focus on habitat quality, the patch density is kept constant to maintain a stable landscape structure throughout the simulations (see Supplement S2 for the effect of patch density) and we alter only habitat quality for each landscape element (with a 5 interval, from 1 to 99, respectively hostile to

favourable). The landscape is a 500 pixels wide square treated as a torus, and is composed of 10 to 12 fields surrounded by 4 pixels-wide hedgerows and 5 pixels-wide GFMs (similar to a typical bocage landscape, Burel et al., 1998; Thenail and Baudry, 2004). Although the field-GFM-hedgerow trio is used as an example for clarity, the structure could apply to fields surrounded by other types of borders.

S1.2.7.2 RDM Parameters

The values used in our case study for foraging parameters are provided in Table 1. They are designed to represent two hypothetical species, to illustrate the sensitivity of the model to differences in habitat sensitivity. An "insensitive species" with a movement behaviour that allows individuals to free themselves from local habitat conditions to reach more easily another region of the landscape: in the RDM model that species is characterised by a lower value of the sensitivity to habitat quality parameter (Table1). On the contrary, the "sensitive species" is characterised by a movement behaviour that depended more strongly on local conditions (Table 1) with a higher sensitivity to habitat and lower directional persistence.

S2 Effects of movement ability, population size and patch density

Mean field residence time

Proportion of unvisited pixels

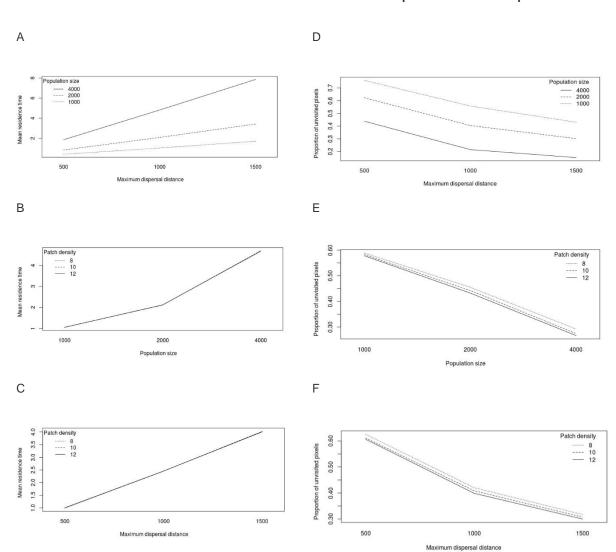


Figure S2.1. Mean field residence time (in ts, A — C) and proportion of unvisited field pixels ($0 \ge p \le 1$, D-F) as a function of interactions between movement ability (pixels, A, C, D, F), population size (number of individuals at initiation, A, B, D, E), and patch density (number of fields in the landscape, B, C, E, F). C: the three lines are shown but overlap.

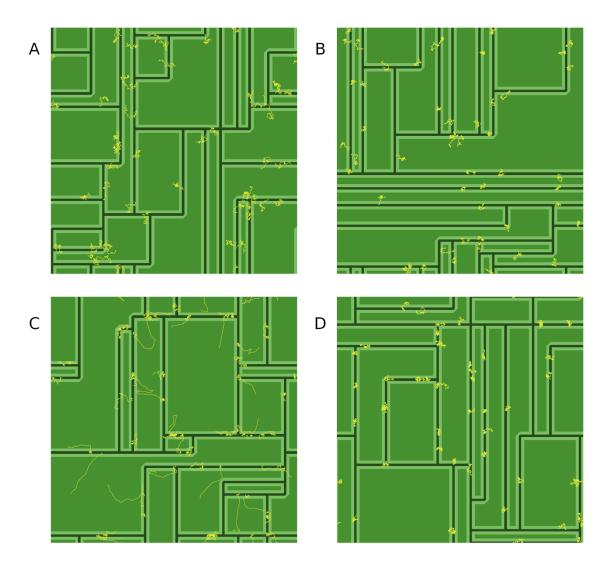


Figure S3.1. Examples of individual movement paths (yellow lines) generated by the RDM model (Routine & Direct Moves: A, C) and the SMS (Stochastic Movement Simulator: B, D). Agricultural fields (dark green) are separated by grassy field margins (light green) surrounding hedgerows (black). The habitat qualities of agricultural fields, grassy field margins and hedgerows vary (respectively of quality 55, 30, 15 in figures 5A, B and 15, 30, 55 in figures 5C, D). N = 50 individual paths, paths length = 200 p.

References

Altieri, M. A. and Policy, S.: The Nature and F unction of Biodiversity in Agriculture, Ecol., 5–8, 2001.

Arrignon, F., Deconchat, M., Sarthou, J. P., Balent, G. and Monteil, C.: Modelling the overwintering strategy of a beneficial insect in a heterogeneous landscape using a multi-agent system, Ecol. Modell., 205(3–4), 423–436, doi:10.1016/j.ecolmodel.2007.03.006, 2007.

Aviron, S., Kindlmann, P. and Burel, F.: Conservation of butterfly populations in dynamic landscapes: The role of farming practices and landscape mosaic, Ecol. Modell., 205(1–2), 135–145, 2007.

Baguette, M. and Van Dyck, H.: Landscape connectivity and animal behaviour: functional grain as a key

- 243 determinant for dispersal, Landsc. Ecol., 22(8), 1117–1129, doi:10.1007/s10980-007-9108-4, 2007.
- 244 Barbosa, P.: Conservation biological control, Academic Press., 1998.
- 245 Bartumeus, F., da Luz, M. G. E., Viswanathan, G. M. and Catalan, J.: Animal Search Strategies: a Quantitative
- 246 Random-Walk Analysis, Ecology, 86(11), 3078–3087, doi:10.1890/04-1806, 2005.
- 247 Bell, W. J.: Searching behaviour. The behavioural ecology of finding resources., Chapman & Hall, London., 1991.
- 248 Bianchi, F. J. J. A., Booij, C. J. H. and Tscharntke, T.: Sustainable pest regulation in agricultural landscapes: A
- review on landscape composition, biodiversity and natural pest control, Proc. R. Soc. B Biol. Sci., 273(1595),
- 250 1715–1727, doi:10.1098/rspb.2006.3530, 2006.
- 251 Bocedi, G., Palmer, S. C. F., Pe'er, G., Heikkinen, R. K., Matsinos, Y. G., Watts, K. and Travis, J. M. J.:
- 252 RangeShifter: a platform for modelling spatial eco-evolutionary dynamics and species' responses to
- environmental changes, edited by R. Freckleton, Methods Ecol. Evol., 5(4), 388-396, doi:10.1111/2041-
- 254 210X.12162, 2014.
- Bonte, D., Van Dyck, H., Bullock, J. M., Coulon, A., Delgado, M., Gibbs, M., Lehouck, V., Matthysen, E., Mustin,
- 256 K., Saastamoinen, M., Schtickzelle, N., Stevens, V. M., Vandewoestijne, S., Baguette, M., Barton, K., Benton, T.
- 257 G., Chaput-Bardy, A., Clobert, J., Dytham, C., Hovestadt, T., Meier, C. M., Palmer, S. C. F., Turlure, C. and
- 258 Travis, J. M. J.: Costs of dispersal., Biol. Rev. Camb. Philos. Soc., doi:10.1111/j.1469-185X.2011.00201.x, 2012.
- 259 Börger, L., Dalziel, B. D. and Fryxell, J. M.: Are there general mechanisms of animal home range behaviour? A
- 260 review and prospects for future research, Ecol. Lett., 11(6), 637–650, doi:10.1111/j.1461-0248.2008.01182.x,
- 261 2008.
- 262 Broekhuizen, S., C.A., H., van't Maaskamp, F. and Pauwels, T.: The importance of hedgerows for leading
- 263 migrating badgers Meles meles, Lutra, 29, 54–65, 1986.
- Burel, F., Baudry, J., Butet, A., Clergeau, P., Delettre, Y., Le Coeur, D., Dubs, F., Morvan, N., Paillat, G., Petit, S.,
- 265 Thenail, C., Brunel, E. and Lefeuvre, J. C.: Comparative biodiversity along a gradient of agricultural landscapes,
- Acta Oecologica-International J. Ecol., 19(1), 47–60, 1998.
- 267 Clobert, J., Danchin, E., Dhondt, A. A. and Nichols, J. D.: Dispersal, Oxford University Press, Oxford., 2001.
- 268 Codling, E. A., Plank, M. J. and Benhamou, S.: Random walk models in biology, J. R. Soc. Interface, 5(25), 813-
- 269 834, doi:10.1098/rsif.2008.0014, 2008.
- 270 Conradt, L., Zollner, P. A., Roper, T. J., Frank, K. and Thomas, C. D.: Foray search: An effective systematic
- 271 dispersal strategy in fragmented landscapes, Am. Nat., 161(6), 905–915, 2003.
- 272 Cook, S. M., Khan, Z. R. and Pickett, J. A.: The Use of Push-Pull Strategies in Integrated Pest Management,
- 273 Annu. Rev. Entomol., 52(1), 375–400, doi:10.1146/annurev.ento.52.110405.091407, 2007.
- Coombes, D. S. and Sothertons, N. W.: The dispersal and distribution of polyphagous predatory Coleoptera in
- 275 cereals, Ann. Appl. Biol., 108(3), 461-474, doi:10.1111/j.1744-7348.1986.tb01985.x, 1986.
- Coulon, A., Aben, J., Palmer, S. C. F., Stevens, V. M., Callens, T., Strubbe, D., Lens, L., Matthysen, E., Baguette,
- 277 M. and Travis, J. M. J.: A stochastic movement simulator improves estimates of landscape connectivity, Ecology,
- 278 96(8), 2203–2213, doi:10.1890/14-1690.1, 2015.
- 279 Delattre, T., Burel, F., Humeau, A., Stevens, V. M., Vernon, P. and Baguette, M.: Dispersal mood revealed by

- shifts from routine to direct flights in the meadow brown butterfly Maniola jurtina, Oikos, 119(12), 1900–1908,
- 281 doi:10.1111/j.1600-0706.2010.18615.x, 2010a.
- 282 Delattre, T., Pichancourt, J.-B., Burel, F. and Kindlmann, P.: Grassy field margins as potential corridors for
- 283 butterflies in agricultural landscapes: A simulation study, Ecol. Modell., 221(2), 370-377,
- 284 doi:10.1016/j.ecolmodel.2009.10.010, 2010b.
- Delattre, T., Vernon, P. and Burel, F.: An Agri-environmental scheme enhances butterfly dispersal in European
- agricultural landscapes, Agric. Ecosyst. Environ., 166, 102–109, 2013.
- 287 Dennis, P. and Fry, G. L. A.: Field margins: can they enhance natural enemy population densities and general
- arthropod diversity on farmland?, Agric. Ecosyst. Environ., 40(1–4), 95–115, 1992.
- 289 Denys, C. and Tscharntke, T.: Plant-insect communities and predator-prey ratios in field margin strips, adjacent
- 290 crop fields, and fallows, Oecologia, 130(2), 315–324, doi:10.1007/s004420100796, 2002.
- 291 Dover, J. W. and Fry, G. L. A.: Experimental simulation of some visual and physical components of a hedge and
- the effects on butterfly behaviour in an agricultural landscape, Entomol. Exp. Appl., 100(2), 221–233, 2001.
- 293 Dunning, J., Danielson, B. and Pulliam, H.: Ecological processes that affect populations in complex landscapes,
- 294 Oikos, 1992.
- 295 Van Dyck, H. and Baguette, M.: Dispersal behaviour in fragmented landscapes: Routine or special movements?,
- 296 Basic Appl. Ecol., 6(6), 535–545, doi:10.1016/j.baae.2005.03.005, 2005.
- 297 Ernoult, A., Vialatte, A., Butet, A., Michel, N., Rantier, Y., Jambon, O. and Burel, F.: Grassy strips in their
- 298 landscape context, their role as new habitat for biodiversity, Agric. Ecosyst. Environ., 166, 15-27,
- 299 doi:10.1016/j.agee.2012.07.004, 2013.
- Fahrig, L., Baudry, J., Brotons, L., Burel, F. G., Crist, T. O., Fuller, R. J., Sirami, C., Siriwardena, G. M. and
- 301 Martin, J.-L.: Functional landscape heterogeneity and animal biodiversity in agricultural landscapes, Ecol. Lett.,
- 302 14(2), 101–112, doi:10.1111/j.1461-0248.2010.01559.x, 2010.
- 303 Griffiths, G. J. K., Holland, J. M., Bailey, A. and Thomas, M. B.: Efficacy and economics of shelter habitats for
- 304 conservation biological control, Biol. Control, 45(2), 200–209, doi:10.1016/j.biocontrol.2007.09.002, 2008.
- 305 Grimm, V. and Railsback, S. F.: Individual-based Modeling and Ecology, Princeton University Press, Princeton.,
- 306 2005.
- 307 Grimm, V., Berger, U., Bastiansen, F., Eliassen, S., Ginot, V., Giske, J., Goss-Custard, J., Grand, T., Heinz, S. K.
- and Huse, G.: A standard protocol for describing individual-based and agent-based models, Ecol. Modell., 198(1–
- 309 2), 115-126, doi:10.1016/j.ecolmodel.2006.04.023, 2006.
- Grimm, V., Berger, U., DeAngelis, D. L., Polhill, J. G., Giske, J. and Railsback, S. F.: The ODD protocol: A review
- and first update, Ecol. Modell., 221(23), 2760–2768, doi:10.1016/j.ecolmodel.2010.08.019, 2010.
- 312 Haaland, C. and Bersier, L.-F.: What can sown wildflower strips contribute to butterfly conservation?: an example
- 313 from a Swiss lowland agricultural landscape, J. Insect Conserv., 1–9, doi:10.1007/s10841-010-9353-8, 2010.
- 314 Haddad: Corridor use predicted from behaviors at habitat boundaries, Am. Nat., 153(2), 215-227,
- 315 doi:10.2307/2463582, 1999.
- 316 Hamilton, N.: ggtern: An Extension to "ggplot2", for the Creation of Ternary Diagrams, 2017.

- 317 Al Hassan, D., Georgelin, E., Delattre, T., Burel, F., Plantegenest, M., Kindlmann, P. and Butet, A.: Does the
- 318 presence of grassy strips and landscape grain affect the spatial distribution of aphids and their carabid
- 319 predators?, Agric. For. Entomol., 15(1), 24–33, 2013.
- 320 Human Rights Council: Report of the Special Rapporteur on the right to food., 2017.
- 321 Kindlmann, P., Aviron, S., Burel, F. and Ouin, A.: Can the assumption of a non-random search improve our
- 322 prediction of butterfly fluxes between resource patches?, Ecol. Entomol., 29(4), 447-456, doi:10.1111/j.0307-
- 323 6946.2004.00614.x, 2004.
- 324 Kindlmann, P., Aviron, S. and Burel, F.: When is landscape matrix important for determining animal fluxes
- 325 between resource patches?, Ecol. Complex., 2(2), 150–158, doi:10.1016/j.ecocom.2004.11.007, 2005.
- 326 Kleijn, D., Baquero, R. A., Clough, Y., Díaz, M., De Esteban, J., Fernández, F., Gabriel, D., Herzog, F.,
- 327 Holzschuh, A., Jöhl, R., Knop, E., Kruess, A., Marshall, E. J. P., Steffan-Dewenter, I., Tscharntke, T., Verhulst, J.,
- West, T. M. and Yela, J. L.: Mixed biodiversity benefits of agri-environment schemes in five European countries,
- 329 Ecol. Lett., 9(3), 243–254, doi:10.1111/j.1461-0248.2005.00869.x, 2006.
- 330 Kremen, C., Williams, N. M., Aizen, M. A., Gemmill-Herren, B., LeBuhn, G., Minckley, R., Packer, L., Potts, S. G.,
- 331 Roulston, T., Steffan-Dwenter, I., Vasquez, D. P., Winfree, R., Adams, L., Crone, E. E., Greenleaf, S. S., Keitt, T.
- 332 H., Klein, A. M., Regetez, J. and Ricketts, T. H.: Pollination and other ecosystem services produced by mobile
- organisms: a conceptual framework for the effects of land-use change, Ecol. Lett., 10, 299–314, 2007.
- Long, S. J.: Regression models for categorical and limited dependent variables, Sage., 1997.
- 335 Lys, J.-A., Zimmermann, M. and Nentwig, W.: Increase in activity density and species number of carabid beetles
- 336 in cereals as a result of strip-management, Entomol. Exp. Appl., 73(1), 1-9, doi:10.1111/j.1570-
- 337 7458.1994.tb01833.x, 1994.
- Mailloux, J., Le Bellec, F., Kreiter, S., Tixier, M.-S. and Dubois, P.: Influence of ground cover management on
- 339 diversity and density of phytoseiid mites (Acari: Phytoseiidae) in Guadeloupean citrus orchards, Exp. Appl.
- 340 Acarol., 52(3), 275–290, doi:10.1007/s10493-010-9367-7, 2010.
- 341 Malézieux, E., Crozat, Y., Dupraz, C., Laurans, M., Makowski, D., Ozier-Lafontaine, H., Rapidel, B., De
- 342 Tourdonnet, S. and Valantin-Morison, M.: Mixing plant species in cropping systems: Concepts, tools and models:
- A review, in Sustainable Agriculture, pp. 329–353, Springer Netherlands, Dordrecht., 2009.
- 344 Mitchell, M. S. and Powell, R. A.: A mechanistic home range model for optimal use of spatially distributed
- 345 resources, Ecol. Modell., 177(1–2), 209–232, doi:10.1016/j.ecolmodel.2004.01.015, 2004.
- 346 Olson, D. M. and Wäckers, F. L.: Management of field margins to maximize multiple ecological services, J. Appl.
- 347 Ecol., 44(1), 13-21, 2007.
- 348 Palmer, S. C. F., Coulon, A. and Travis, J. M. J.: Introducing a "stochastic movement simulator" for estimating
- 349 habitat connectivity, Methods Ecol. Evol., 2(3), 258–268, doi:10.1111/j.2041-210X.2010.00073.x, 2011.
- Papaïx, J., Touzeau, S., Monod, H. and Lannou, C.: Can epidemic control be achieved by altering landscape
- 351 connectivity in agricultural systems?, Ecol. Modell., 284, 35–47, doi:10.1016/j.ecolmodel.2014.04.014, 2014.
- 352 Pulido, F. and Berthold, P.: Current selection for lower migratory activity will drive the evolution of residency in a
- 353 migratory bird population., Proc. Natl. Acad. Sci. U. S. A., 107(16), 7341–7346, doi:10.1073/pnas.0910361107,
- 354 2010.

- R Core Team: R: a Language and Environment for Statistical computing, , doi:10.1007/978-3-540-74686-7, 2011.
- 356 Ratnadass, A., Fernandes, P., Avelino, J. and Habib, R.: Plant species diversity for sustainable management of
- 357 crop pests and diseases in agroecosystems: a review, Agron. Sustain. Dev., 32(1), 273-303,
- 358 doi:10.1007/s13593-011-0022-4, 2012.
- 359 Root-Bernstein, M. and Jaksic, F. M.: Making research relevant? Ecological methods and the ecosystem services
- 360 framework, Earth's Futur., 5(7), 664–678, doi:10.1002/2016EF000501, 2017.
- 361 Schellhorn, N. A. and Sork, V. L.: The impact of weed diversity on insect population dynamics and crop yield in
- 362 collards, Brassica oleraceae (Brassicaceae), Oecologia, 111(2), 233-240, doi:10.1007/s004420050230, 1997.
- 363 Shelton, A. M. and Badenes-Pérez, F. R.: Concepts and applications of trap cropping in pest management, Annu.
- 364 Rev. Entomol., 51(1), 285–308, doi:10.1146/annurev.ento.51.110104.150959, 2006.
- 365 Simberloff, D. and Stiling, P.: Risks of species introduced for biological control, Biol. Conserv., 78(1–2), 185–192,
- 366 doi:10.1016/0006-3207(96)00027-4, 1996.
- 367 Stevens, V. M., Wesselingh, R. A. and Baguette, M.: Demographic processes in a small, isolated population of
- natterjack toads (Bufo calamita) in southern Belgium, Herpetol. J., 13(2), 59-67, 2003.
- 369 Thenail, C. and Baudry, J.: Variation of farm spatial land use pattern according to the structure of the hedgerow
- 370 network (bocage) landscape: a case study in northeast Brittany, Agric. Ecosyst. Environ., 101(1), 53-72,
- 371 doi:10.1016/s0167-8809(03)00199-3, 2004.
- Theunissen, J. and Schelling, G.: Pest and disease management by intercropping: Suppression of thrips and rust
- in leek, Int. J. Pest Manag., 42(4), 227–234, doi:10.1080/09670879609372000, 1996.
- Tisue, S. and Wilensky, U.: Netlogo: A simple environment for modeling complexity, vol. 21, pp. 16–21, Boston,
- 375 MA., 2004.
- 376 Tylianakis, J. M., Didham, R. K. and Wratten, S. D.: Improved fitness of aphid parasitoids receiving resource
- 377 subsidies, Ecology, 85(3), 658–666, doi:10.1890/03-0222, 2004.
- 378 Vasseur, C., Joannon, A., Aviron, S., Burel, F., Meynard, J.-M. and Baudry, J.: The cropping systems mosaic:
- 379 How does the hidden heterogeneity of agricultural landscapes drive arthropod populations?, Agric. Ecosyst.
- 380 Environ., 166, 3–14, 2013.
- 381 Veres, A., Petit, S., Conord, C. and Lavigne, C.: Does landscape composition affect pest abundance and their
- 382 control by natural enemies? A review, Agric. Ecosyst. Environ., 166, 110–117, doi:10.1016/J.AGEE.2011.05.027,
- 383 2013.

- Wäckers, F. L., van Rijn, P. C. J. and Bruin, J.: Plant-Provided Food for Carnivorous Insects, edited by Felix L.
- Wäckers, Paul C. J. van Rijn, and Jan Bruin, Cambridge University Press., 2005.

13