

Applying seed germination studies in fire management for biodiversity conservation in south-eastern Australia

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We examine the patterns of germination response to fire in the fire-prone flora of the Sydney basin, south-eastern Australia, using examples from several decades of research. The flora shows a strong response to fire-related germination cues. Most species show an interaction between heat and smoke, a number respond only to heat, whilst a few are likely to respond only to smoke. Many recruit in the first 12 months after fire and show no obvious seasonal patterns of recruitment, whilst several species have a strong seasonal germination requirement, even in this essentially aseasonal rainfall region. Key challenges remaining include designing future seed germination studies within the context of informing the germination response surface to smoke and heat interactions, and incorporation of the impact of varying soil moisture on seed germination post-fire, including its affect on resetting of seed dormancy. An understanding of the resilience of species to frequent fire also requires further work, to identify species and functional types most at risk. This work must ideally be integrated within the framework of the management of fire regimes that will change under a changing climate. We suggest that the functional classification of plant types in relation to fire could be enhanced by a consideration of both the type of germination response to fire (type of cues required) and the timing of the response (seasonally driven in response to seed dormancy characteristics, or independent of season). We provide a simplified version of such an addition to functional trait classification in relation to fire.

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Fire management is a complex interaction of decisions incorporating the major elements of protection of human life and property and the long-term conservation of biodiversity. The management of biodiversity in fire-prone areas is dependent upon the understanding of how components of the fire regime interact and impact on plants and animals. For plants, much effort has gone into developing simple fire response classifications (Gill 1981, Gill and Bradstock 1992, Pausas 1999, Pausas and Lavorel 2003, Pausas et al. 2004) and gaining an understanding of how critical components of the fire regime, such as fire frequen-

cy, may influence species persistence (Gill and McCarthy 1998, Whelan 2002, Pausas and Bradstock 2007). However, the understanding of the significant component of species in fire-prone floras that have a soil-stored seed bank is limited, and may be poorly addressed in simple fire-response classifications. The conservation management of this large group of species in relation to fire has usually been dealt with simply in terms of whether or not species have an ability to resprout (seeders versus sprouters) (Gill and Bradstock 1992). The nature of propagule persistence (Pausas et al. 2004) and the application of seed bank types

(canopy, soil persistent, soil transient) (Keith et al. 2007) have also been considered. For both sprouters and seeders, the impact of fires and the fire regime on recruitment will have a major influence on persistence in the long term, as seed germination is often cued to fire (Keeley 1987, Auld and O'Connell 1991, Bell 1999, Paula and Pausas 2008). Consequently, understanding how components of the fire regime influence recruitment from seed is a key issue for fire management and biodiversity conservation.

The role of fire and seed germination has largely been derived from laboratory studies and observations of post-fire flushes of germination in the field. Laboratory studies on how fire may affect seed germination have focussed on germination cues such as heat and smoke and their interaction with seed dormancy characteristics (Keeley 1987, Auld and O'Connell 1991, Brown 1993, Dixon et al. 1995, Keeley and Fotheringham 1998, Ooi et al. 2006, Merritt et al. 2007, Ooi 2007, Thomas et al. 2007) or resprouting capacity (Paula and Pausas 2008). Most studies have limited replication across the fire cues and often do not examine interactive effects (Thomas et al. 2007). In this context we need to consider the response surface (sensu Thomas et al. 2007) that represents the germination response of seeds over the full range of temperatures, durations of heating and smoking, and smoke concentrations likely to be experienced in the field. Currently most studies only address a small subset of this response surface and a synthesis of the implications of such studies may be limited by the lack of understanding of the germination across the full response surface. Field studies have clearly indicated strong post-fire germination in many taxa in fire-prone habitats (Tyler 1995, 1996, Auld and Tozer 1995, Keith et al. 2002). What is more limited is a clearer understanding of the relative seasonal timing of emergence (cf. Ooi et al. 2004, Ooi 2007) and the proportion of the seed bank that emerges versus the proportion that persists in the soil (Auld and Denham 2006).

In this paper, we examine data from over two decades of intensive research into fire-related germination cues of species from the Sydney basin region in south-eastern Australia. We: 1) ask whether existing studies can be used to inform management for biodiversity conservation in a fire-prone region which has strong competing needs for the protection of human life and property, 2) examine how such studies may have influenced fire management in fire-prone habitats in eastern Australia and where further refinement may be necessary, and 3) highlight those areas that urgently require future research to aid conservation management.

The Sydney basin flora

We collected information from a broad range of studies, summary papers and reports that addressed the impacts of fire on germination in plants from fire-prone habitats

of the Sydney basin, south-eastern Australia. This area is a major fire-prone region in eastern Australia and includes the urban metropolis of Sydney with a population of over 5 million people and associated infrastructure. This information was integrated with current fire management practices in this area.

Within the Sydney basin, there is a rich and diverse flora. There are some 2000 plant species that occur in fire-prone habitats including a number of endemic species and species considered to be threatened. The vegetation at any one location is made up of a mixture of sprouter and seeder species. Approximately 10% of all species have a canopy seed bank (including *Banksia*, *Hakea*, *Allocasuarina* and *Callitris*). In the order of 15% of species (although this figure may be higher) have no seed bank at the time of a fire, instead they resprout after fire and then flower and produce a transient seed bank (often called obligate resprouters). These contain a number of members of the Liliaceae (sensu lato) as well as some shrubby species of *Angophora*, *Lomatia* and *Telopea*. The great majority of species in fire-prone habitats (approximately 75%) have soil seed banks with some level of seed persistence. In these species there is a germination response in relation to fire cues. For the obligate seeders this is necessary for population persistence post-fire, while for the facultative resprouters successful recruitment from seeds may not be necessary after each fire, given the ability of adult plants to regrow after fire. The proportion of sprouters/seeders in the Sydney basin region is around 0.52/0.48, although in many other fire-prone regions the obligate seeding proportion is generally much lower (Pausas et al. 2004). In the Sydney basin, species with persistent soil seed banks belong to a range of plant families, and examples include Fabaceae (*Acacia*, *Dillwynia* and *Pultenaea*), Rutaceae (*Boronia*, *Philotheca*), Ericaceae (*Epacris*, *Leucopogon*), Myrtaceae (*Baeckea*, *Kunzea*), Poaceae (*Eragrostis*, *Microlaena*) and Proteaceae (*Grevillea*, *Persoonia*).

Managing fire frequency

Fire in the Sydney basin region is influenced by a variety of factors, including the extensive reserve network of flammable vegetation on poor nutrient soils, past clearing, weed encroachment (particularly where there has been nutrient enrichment near urban centres, Thompson and Leishman 2005) and the impact of a changing climate. Fire management for biodiversity conservation has concentrated on two aspects based on available data. Firstly, species have been classified according to their fire response (Gill and Bradstock 1992). Secondly, high fire frequency has been listed in the state of New South Wales (NSW) as a key threatening process (NSW Scientific Committee 2000) and, in response, fire-management attention has focussed on identifying minimal fire-free intervals that are required by species and assemblages of species. One such example

is the genus *Persoonia* (Proteaceae). This group contains approximately 35 taxa in the Sydney basin (Weston 2002), of which some 76% are seeders. A number of these seeders are listed as threatened (Table 1). Where data are available, time to first flowering (primary juvenile period) may take 3–10 years (Benson 1985, Table 1), well beyond many co-habiting species (Keith et al. 2002). However, data for many species are lacking (Table 1). For most species, there is little flowering at the time of first maturity (Auld et al. 2007) and it takes several more years for peak seed production to be reached. For several species, this means that most seed production occurs at an age double that of the primary juvenile period (NSW National Parks and Wildlife Service 2000, Auld et al. 2007). The average fire return interval for these habitats is 7–17 years (Bradstock and Kenny 2003) and fires may occur at a frequency as high as every 5 years, particularly where arson or fire management for protection of human life and property oc-

cur. Fire management for threatened *Persoonia* species has consequently set minimum fire return intervals of over a decade (Table 1) in order to reduce the risk of fire-driven population declines or extinctions.

Seed germination cues and fire

Germination studies in south-eastern Australia have focussed primarily on the impacts of heat and smoke alone, and more recently on the interaction of these two factors. Few studies have attempted to examine the full germination response surface across the breadth of the fire cues, with most simply testing a small component of it, i.e. a single temperature and duration of heating, or a single smoke concentration. Three groupings of species responses are apparent. Firstly, there are a number of predominately hard seeded species where the degree of heating controls

Table 1. Fire response and time to maturity in *Persoonia* species from the Sydney basin (data compiled from Benson and McDougall 2000, NSW National Parks and Wildlife Service fire response register 2002, T. D. Auld unpubl.). Threat status: en – endangered, vu – vulnerable. The minimum fire-free interval is the recommended interval between fires that must be allowed for species persistence (from NSW Rural Fire Service 2006, T. D. Auld unpubl.).

Species	Fire response	Threat status	Primary juvenile period (years)	Minimum fire-free interval (years)
<i>P. acerosa</i>	seeder	vu	?	10+
<i>P. acuminata</i>	?	not threatened	?	?
<i>P. bargoensis</i>	seeder	en	?	10+
<i>P. chamaepeuce</i>	sprouter	unknown	?	?
<i>P. chamaepitys</i>	seeder	not threatened	?	?
<i>P. glaucescens</i>	seeder	en	?	15+
<i>P. hindii</i>	sprouter	en	?	10+
<i>P. hirsuta</i>	seeder	en	?	10+
<i>P. isophylla</i>	seeder	not threatened	3–7+	10+
<i>P. lanceolata</i>	seeder	not threatened	3–7+	10+
<i>P. laurina</i>	sprouter	unknown	?	?
<i>P. laxa</i>	seeder	extinct	?	?
<i>P. levis</i>	sprouter	not threatened	7+	10+
<i>P. linearis</i>	sprouter	not threatened	?	?
<i>P. marginata</i>	sprouter	vu	?	10+
<i>P. microphylla</i>	?	unknown	?	?
<i>P. mollis</i>	seeder	en (subspecies)	8–10	12+
<i>P. myrtilloides</i>	seeder	unknown	?	?
<i>P. nutans</i>	seeder	en	?	10+
<i>P. oblongata</i>	seeder	enknown	?	?
<i>P. oxycoccooides</i>	sprouter	rare	?	?
<i>P. pauciflora</i>	seeder	en	?	no fire
<i>P. pinifolia</i>	seeder	not threatened	5–8+	12+
<i>P. recedens</i>	?	rare	?	?
<i>P. rigida</i>	seeder	unknown	?	?

Table 2. Incorporation of fire related seed germination strategies into a functional classification of plants. This level of classification would sit below the distinction between persistent and transient seed banks under each of the seeder and sprouter categories.

Seed bank type	Fire response cues	Seasonal fire response	Comments/examples from Sydney basin
Soil transient	none	no	common, liliaceae sensu lato <i>Angophora</i> , <i>Doryanthes</i> , <i>Lomatia</i> , <i>Telopea</i> , <i>Xanthorrhoea</i>
		yes	unknown
Soil persistent	heat only	no	common, Fabaceae
		yes	rare/unknown
	smoke only	no	uncommon
		yes	currently unknown but likely
	heat and smoke interaction	no	common and widespread, e.g. <i>Grevillea</i> , <i>Epacris</i> . Likely in <i>Persoonia</i> .
	yes	occasional, e.g. <i>Leucopogon</i> , <i>Kunzea</i>	

the breaking of seed dormancy. In the Sydney basin, this represents some 250 species (12–15% of total flora) and includes most legumes (Fabaceae) and *Dodonaea* species (Sapindaceae) (Floyd 1966, 1976, Auld and O’Connell 1991, Keith 1996, Ooi 2007; Table 2). For these species, during a fire, the magnitude and duration of soil heating will decline with soil depth (Fig. 1). This is because soil heating is influenced by the consumption of fine fuels at the soil surface during the passage of a fire (Bradstock and Auld 1995) and the soil has insulating properties. Secondly, there has been a more limited examination of permeable seeds and seeds with physiological dormancy. Such species often have germination enhanced by smoke or the combination of heat and smoke (e.g. Ericaceae, Rutaceae and a number of small seeded Myrtaceae, e.g. *Kunzea*). Whilst only a small proportion (<10% of the total flora) are thought to respond to smoke alone (Kenny 2003), we estimate that approximately 900 species (25–45% of the total flora) have complex smoke × heat interactions (Keith 1997, Kenny 2000, 2003, Morris 2000, Thomas et al. 2003, 2007, Ooi et al. 2004, Clarke and French 2005). For these species, the relationship between fire-cues and soil depth is less certain. A depth related response is expected for species with interactions between heat and smoke (due to declines in heating with increasing soil depth). For those species with germination enhanced by a smoke effect only, the smoke cue may operate independently of soil depth or have some other relationship with soil depth (Fig. 1), although this remains to be explored.

Finally, there are a number of taxa that currently show no response to fire cues, or have a response that is essentially unknown. Much of this is primarily because of the limited understanding of physiological dormancy and fire cues. Although dormancy has not been precisely classified

for many taxa, it is very likely that more than 200 species (10% of the total flora, e.g. *Hibbertia*, *Persoonia*, Rutaceae, Apiaceae, many Ericaceae, Ooi et al. 2004, Ooi 2007) have a physiological dormancy component. For many of these taxa, a response to either heat or smoke alone has been documented but is often variable, and the response is often not an obligate one. Additionally, as for many other species, the full response surface has not been examined. We consider it likely that the taxa which are difficult to germinate from seed, or have poorly known germination

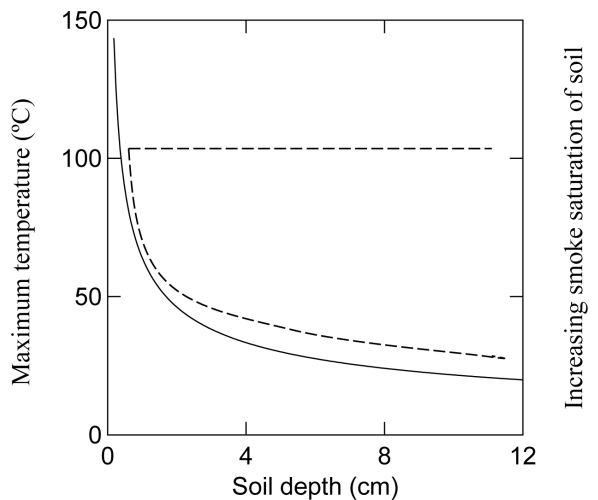


Fig. 1. Possible variation of fire cues with soil depth. Solid line – heat profile from a typical wildfire (T. D. Auld unpubl.). The dashed lines represent a range of possible relationships between smoke concentration and soil depth, varying from complete saturation at all depths (horizontal line) to a rapid decline in saturation with soil depth.

mechanisms, are those that have a physiological dormancy component (e.g. physiological, morphophysiological or physical + physiological dormancy classes), as is expected for many difficult to germinate species from Western Australia (Merritt et al. 2007).

Seasonal impacts on germination

For a number of taxa (some 15–30% of the flora) there are no seasonal requirements for germination (Table 2), independent of the fire cue that promotes germination. Most Fabaceae falls into this group. Unlike fire-prone habitats in Mediterranean regions, rainfall is more or less evenly spread across seasons in the Sydney basin. As a consequence, fires and opportunities for post-fire recruitment can occur at any time in this region. In contrast, a few species are known to only germinate in a certain season (Table 2). They have specific temperature needs for either embryo maturation and/or germination e.g. all *Leucopogon* species (Ericaceae) (Ooi et al. 2004). Like *Leucopogon* spp., a number of species have a strong seasonal temperature re-

quirement for germination (e.g. *Kunzea* spp., *Baeckea* spp., *Epacris* spp., *Drachophyllum* spp., *Goodenia* spp., Thomas 2004). In fire-prone Mediterranean systems, this seasonal response is expected to be the predominant functional type, given the lack of available moisture in summer.

In addition, whilst fire may promote germination, dormancy may need to be broken initially by seasonal temperatures. A seasonal cue could effectively reset dormancy or prevent germination outside a particular seasonal window. This seasonal requirement can override the fire-related germination cues, e.g. *Kunzea* spp. (Auld et al. 2000, Thomas 2004, Ooi 2007, Merritt et al. 2007) and, as a consequence, the germination of species with strong seasonal requirements may be delayed post-fire relative to those species with no seasonal requirement (Fig. 2). This could give a competitive advantage, under certain fire regimes, to species with no seasonal requirement for germination. The competitive disadvantage in species whose emergence is tied to a season would potentially result from both the delay in seedling emergence, as well as the reduced magnitude of emergence (Fig. 2), e.g. via seed decay or seed predation over the delayed period.

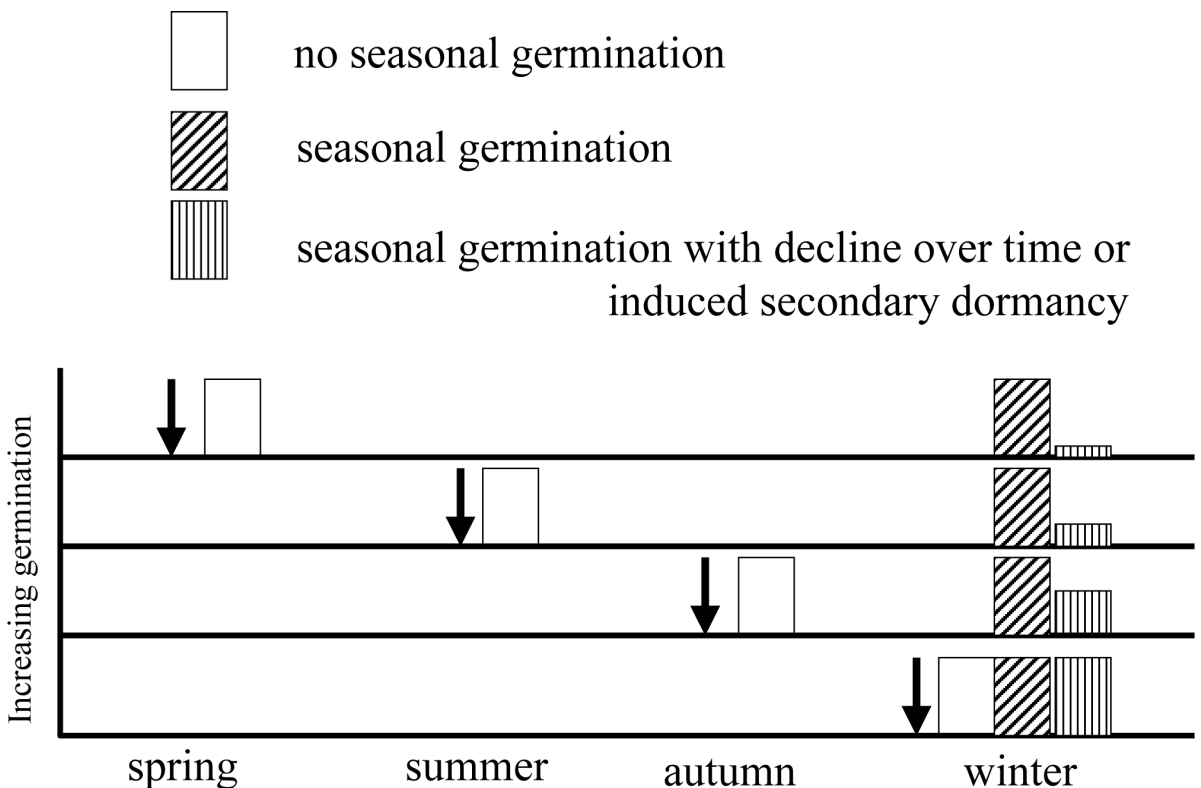


Fig. 2. Impact of season of burn on resultant seedling germination for three classes of species: no seasonal germination impacts (open bars); seasonal germination (sloped hatching); and seasonal germination that declines with the length of time since fire with respect to the favourable season for germination (vertical hatching). Arrows represent four postulated fires (one for each season). The germination level after fires in different seasons is staggered for clarity.

Further research

Most studies of relevance to plant recruitment and fire have either dealt with monitoring the magnitude of species composition after fire or examining germination cues in laboratory studies. These studies illustrate that there is a definite pulse of germination after fire in a broad array of plant species from many different plant families (Keith 1996, Keith et al. 2002). Most recruitment is also confined to soon after fire in non Mediterranean eastern Australian systems (Auld and Tozer 1995, Keith et al. 2002), where rainfall is not strongly seasonal. At the same time, there are some taxa with forms of physiological dormancy that may have delayed emergence, relative to the majority of co-habiting species, with the timing of emergence dependant upon the season of fire (Ooi 2007). More data is needed to determine just how widespread and abundant species with a seasonal germination response are in the wild in this study region.

Laboratory studies have highlighted the importance of both heat and smoke as fire related cues that either break seed dormancy or enhance seed germination. However, our understanding of species responses to fire regimes is still somewhat fragmented and a number of key knowledge gaps remain. Firstly, the development of an understanding of the germination response surface of a species to interactive fire cues needs to be extended. A solid framework is in place (Thomas et al. 2007), but a greater range of taxa need to be examined across the whole potential response surface, as opposed to single temperatures, durations of heating and concentrations of exposure to smoke. As well, the role of variation in soil moisture in influencing seed dormancy and the enhancement of germination (particularly by smoke) remains to be incorporated into this framework. This work then urgently needs to be experimentally tested in the field in experimental fires to determine if the complexity in the laboratory is mirrored in the field. Secondly, the examination of the nature of the resilience of species with soil seed banks to frequent fire is in its infancy. Broad scale generalisations are currently not possible, although there may be patterns that relate directly to seed morphology and the types of fire-related cues that promote germination (Auld and Denham 2006, Auld et al. 2007, Ooi 2007).

Finally, the transfer of practical information to fire managers has essentially concentrated on aspects of fire frequency. Although fire management has applied the concept of variation in fire regimes, therefore addressing fire severity and patchiness to some extent, the significance of other components of the fire regime are still poorly known and little applied. For example, the amount of soil heating during a fire appears to be a good indicator of post-fire germination in hard-seeded (or physically dormant) species and fire managers in eastern Australia have attempted to have variation in local fire severity to encourage recruitment from species that respond to heat. However, fire sea-

son impacts, predicted from laboratory studies (Ooi 2007), are currently largely outside the fire management focus. In addition, fire impacts may be distorted by limiting managed fires to cooler and wetter months when fire can be effectively controlled, even though this may fall outside the historical fire regime (McLoughlin 1998). Given the likely changes to fire regimes under a changing climate (Cary et al. 2006), a more complete understanding of the interaction of components of the fire regimes is warranted.

To this end, it is clear that simple classifications of plants into functional types based on sprouting versus seeding and canopy versus soil seed banks remains of limited use in compiling the complexity of vegetation response to fire. At a minimum, we consider that there is an urgent need to incorporate components of fire-related recruitment into functional classifications. In its simplest form this would incorporate the three basic functional responses identified here (Table 2), i.e. a scaling of fire cues (heat only, smoke only, interaction of heat and smoke) in combination with seed dormancy type (itself a major driver of seasonal responses and the likelihood of resetting dormancy under environmental stresses) and seasonal germination or emergence response.

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